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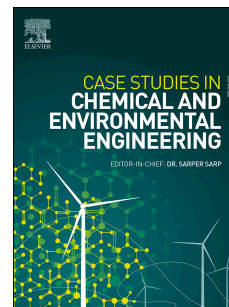


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MouhamedNgounouno Ayiwouo, Luc Leroy MambouNgueyep, Jacques Richard Mache, SifeuTakougang Kingni, Ismaila Ngounouno



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Waters of the Djouzami gold mining site (Adamawa, Cameroon): physicochemical characterization and treatment test by Banasmectite (West, Cameroon).

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Abstract

This paper firstly evaluates the influence of semi-mechanized artisanal gold mining in Djouzami (Adamawa, Cameroon) on the water quality of the site. Samples of runoff, stagnant water and sediments are collected in the study area. These waters are characterized from a physical (pH, electrical conductivity, turbidity and suspended solids) and chemical point of views. The results obtained show that the waters of the semi-mechanized artisanal gold mining site of Djouzami are slightly acid to neutral ($5.93 < \text{pH} < 7.02$), turbid ($46 \text{mg.L}^{-1} < \text{SS} < 200 \text{mg.L}^{-1}$) and polluted by inorganic substances. These include heavy metals such as Cd, Pb and Fe, which exceed the World Health Organization (WHO) standard. The correlation matrix is used to determine relationship existing between the different physicochemical parameters and heavy metals in water and sediments samples. Finally, this paper investigates the ability of Banasmectite (West, Cameroon) denoted BN 61 to remove lead in aqueous medium. It appears that all the lead is adsorbed after adding 0.5 g of clay.

Keywords: Semi-mechanized artisanal gold mining, heavy metals, water, sediments, pollution, smectic clay, adsorption.

1. INTRODUCTION

In Cameroon, there has been more than a decade of increased expansion in small and medium-scale gold mining, particularly in the eastern and Adamawa regions, because of the rising price of the raw material on the world market and the discovery of many deposits. These exploitations are mainly concentrated in the placer and affect the water basins. However, these waters are often too mineralized to be directly usable for drinking water, and in some extreme cases, it is a source of important contamination of the environment [1]. James Lyon said "Water, the victim of mining" thus, the protection of water resources in mining areas becomes imperative for sustainable development. The respect and application of the standards, rules and laws governing the protection of water during mining activities would make it possible not to consider the exploitation of mineral substances as a nightmare for rivers. Physicochemical characterization works has been carried out around the world to evaluate the water quality [2] and specifically show the impact of mining on water resources. These studies were investigated in Morocco in the abandoned Kettara mine [3] and the Zeida mine [4], in Bolivia in the Val Milluni [5], in Ivory Coast in the Bonikro gold mine [6] or in Ghana [7]. They show that this activity also has (during or after) environmental consequences and can become a problem for the human and economic development of society living in this environment. Very few

physicochemical characterization studies of water from mining operations have been investigated in Cameroon. Recently, a study was conducted on the assessment of the surface water quality of the Betare-Oya gold zone in the east region of Cameroon [8]. In the region of Adamawa, no study has been done to the best of the authors knowledge. In rivers, the greatest risk of pollution is related to the presence of heavy metals. Metals are omnipresent in surface waters. However, their concentrations are generally very low, which explains why they are called "trace metals" or "metallic trace elements" (TM). TM are commonly found in the earth's crust. The alteration and erosion of rocks naturally feed surface water into TM [9]. At high concentrations, TM cause adverse effects in humans. As a result, the search for technologies to eliminate these metals in water requires a primordial issue.

During the last decades many researches have focused on adsorption techniques. This technology has already shown its potential in the treatment of inorganic pollutants (heavy metals as Pb (II) and Zn (II)) on water [10]. The adsorption has also the potential to remove toxic organic compounds, natural organic materials and volatile organic compounds. Compared to other technologies, it is an economical technology to remove heavy metals in water. Thus, the research is increasingly oriented towards treatment processes using natural materials less expensive and widely available [11] such as natural zeolites, volcanic ash and especially clays [12,

13]. The application of natural and abundant adsorbents such as clay for the treatment of water is a legitimate way to preserve water capital. Clays remain to this day the natural adsorbent materials which have a high adsorption capacity [14] thanks to their high specific surface, their size and the presence of charge on their surface. For nearly three decades, many research projects have been conducted on clay materials from Cameroon and their potential for recovery. Some authors focused on talc [15], halloysite [16], smectites [17] and kaolinic clays, which nowadays have been the largest clay deposit in Cameroon [18-21]. In the context of valorization of local materials, this study is focus on the use of BanaSmectite (West, Cameroon) as an adsorbent in the process of elimination of heavy metals in aqueous medium.

The main objective of this study is to characterize physically and chemically the waters of the semi-mechanized artisanal gold mining site of Djouzami (Adamawa, Cameroon) and to carry out a test of elimination of lead by adsorption on the smectite of Bana (West, Cameroon). This work is organized as follows: the second section presents materials and methods. The third section presents results and discussions. The last section gives the conclusion which recalls the original results and the main contributions of this study.

2. MATERIALS AND METHODS

The study area is located in Djouzami in the region of Adamawa (Cameroon), Mbere department and the district of Meiganga at around 40 km from the city of Meiganga going to the east of Cameroon. It is precisely located at the east of Djouzami village and bounded by geographical coordinates A (6°14'31.91"N, 14°24'15.00"E), B (6°14'33.53"N, 14°26'22.43"E), C (6°12'36.16"N, 14 ° 24'15.41"E) and D (6°12'36.54"N, 14°26'22.02 "E). This locality belongs to the Mbere ditch. The study area is crossed by the Lom River. The location map of the study area is presented in figure 1.

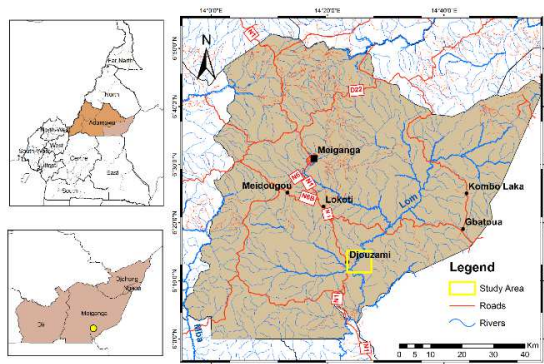
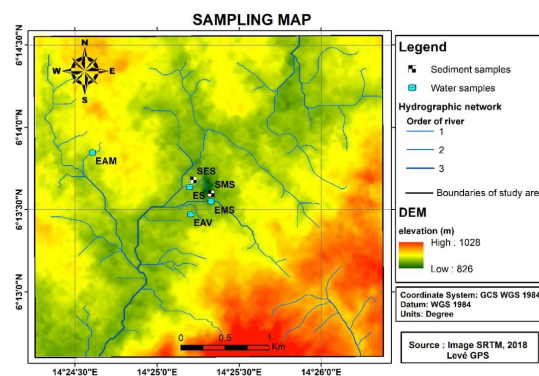


Figure 1: Location map of the study area.

2.1. Sampling, physicochemical parameters of waters and correlation matrix

Four samples of water are taken at different points of the site. **EAM** corresponds to the sampling upstream of the

site on the Lom River; **ES** is the stagnant water; **EMS** corresponds to a sample taken in the middle of the site on the Lom River and **EAV** corresponds to the sample taken downstream of the site. The sampling campaign was done on a sunny day and the flow of the Lom River was relatively average. Water samples are taken from plastic bottles at 30cm below the surface. Sampling of sediment is done at two points on the site. **SMS** corresponds to the sample taken in the middle of the site near Lom and **SES** corresponds to the sediment taken from a pit containing stagnant water. The map sampling of waters and



sediments is presented in figure 2.

Figure 2: Waters and sediments sampling map.

To characterize physically the water samples taken from the mining site, many parameters are determined among which are the pH, the electrical conductivity (EC), the turbidity (Tu) and suspended solids (SS). The pH analysis is carried out in the chemistry Laboratory School of Geology and Mining Engineering of the University of Ngaoundere, Cameroon. Conductivity, turbidity and SS are measured using the analytical protocols of the Chemical Engineering Laboratory of the University Institute of Technology of the University of Ngaoundere, Cameroon. The analysis of heavy metals is carried out at the International Institute of Tropical Agriculture Laboratory of Analysis of Yaounde by using atomic absorption spectrometry (AAS). The apparatus used is the Buck Scientific Atomic Adsorption Spectrometer 205. The analysed metals are copper, zinc, nickel, cadmium, lead and iron. Sediment samples (02) are analysed by portable XRF (Gemius 7000) to determine metal and trace element concentrations.

The correlation matrix used in the work is Pearson's correlation [22].

$$r = \frac{\sum(X-\bar{X})(Y-\bar{Y})}{\sqrt{\sum(X-\bar{X})^2 \sum(Y-\bar{Y})^2}} \quad (1)$$

where, \bar{X} = mean of X variable and \bar{Y} = mean of Y variable.

2.2. Characterization of smectite and its adsorption properties

In this study, the smectic clay is denoted **BN 61**. The analyzes techniques are used to define the physical properties of smectic clay are: Nitrogen adsorption

volumetric analysis to determine the specific surface; determination of the cation exchange capacity (CEC) to estimate the theoretical cation binding power and determination of potential hydrogen (pH) to measure the acidity of a substance. It is determined from the amount of hydrogen ions (H^+) or hydroxide ions (OH^-) contained in the substance. The specific surface area and total pore volume of the sample are determined from nitrogen adsorption and desorption isotherms at 77 K, obtained from a Carlo Erba Sorptomatic 1990 volumetric device, after outgassing the samples overnight at ambient temperature at a pressure lower than 10^{-4} Pa [23]. The specific surface area (SSA) is calculated from adsorption data by applying the Brunauer-Emmet-Teller (B.E.T) method. Cation exchange capacities (CEC) are measured using cobalt hexamine $[Co(NH_3)_6Cl_3]$ as exchangeable ions. The amount of cobalt hexamine fixed by the solid phase is determined using UV-visible spectroscopy. The displaced cations are determined by atomic absorption spectrometry (Perkin-Elmer 1100B). The equilibrium pH of clay suspensions are measured using an analytical radiometer TIM 845. These parameters are measured following the laboratory protocol of the Research Unit Clays, Geochemistry and Sediment Environment (AGEs) of the University of Liege.

The XRD analysis is done on a Bruker AXS model D8 Advanced diffractometer, with Cu-K α radiation, under 40 kV and 30 mA operating conditions, to identify the mineralogical phases of the bulk clay sample. The XRD patterns on the $< 63 \mu m$ fraction are recorded between 2° and 70° using a step scan 0.02° and a step time of 3 seconds. The test consists of saturating the $< 2 \mu m$ clay fractions with 2 M LiCl solution overnight. The Li-exchanged fractions are then rinsed with demineralized water, and oriented aggregates have been prepared on glass slides. The XRD analyses on oriented clay fractions are conducted in sequence on the air-dried slide (N), heated at $300^\circ C$ (H300, 2H), and finally overnight glycerol solvated (G). A d spacing (001) reflection at 9.6 to 10 \AA , is indicative of montmorillonite (octahedral negative layer charge), while a spacing at 16.7 to 17.7 \AA indicates beidellite (tetrahedral negative layer charge).

Chemical analyses are carried out using emission spectrometry. The samples are molded in fused lithium borate ($LiBO_2$) and dissolved in nitric acid. The major elements are determined by Inductive Coupled Plasma by Atomic Emission Spectrometry (ICP-AES), while trace and rare earth elements Inductive Coupled Plasma by Mass Spectrometry (ICP-MS). Relative analytical uncertainties are estimated at 1-5% for major elements except for P_2O_5 (10%). They are up to 5% for most of the trace element concentrations except for Cu (10%). However, uncertainty is high ($>10\%$) for any trace element displaying a low concentration ($<0.1 ppm$). This

analysis is carried out at the *Service d'Analyse des Roches et des Minéraux* (SARM Nancy-France).

The adsorption of lead is studied as a function of the adsorbent mass. The experiments are performed at ambient temperature. In a 125 mL beaker, amount 0.1 to 0.5 g of smectic clay are brought into contact with 50 mL of water containing lead of concentrations $0.089 mg.L^{-1}$ and $0.2 mg.L^{-1}$. The samples are agitated for 1 hour using an IKA RT 5 magnetic stirrer at a temperature of $25^\circ C$ and a speed of 800 rpm.

3. RESULTS AND DISCUSSIONS

The physical analyses of the sampling waters are summarized in Table 1.

Table 1: Physical characteristics of the waters of the artisanal semi-mechanized gold mining site of Djouzami.

Parameters	Samples				WHO standard
	EAM	EMS	ES	EAV	
pH	6.59	6.73	5.93	7.02	$6.5 < pH < 8.5$
EC ($\mu s.cm^{-1}$)	23.3	24.9	7.42	22.4	$14 \mu s.cm^{-1}$
Tu (NTU)	38.23 ± 2.31	235.3 3 ± 4.73	265 ± 4.36	167 \pm 5.29	5 NTU
SS ($mg.L^{-1}$)	46	282	318	200	$50 mg.L^{-1}$

The water samples taken from the mining site have a pH ranging from slightly acidic to neutral. Table 1 shows that the samples EAM, EMS and EAV have a pH which is in agreement with the standard for a surface water in particular $6.5 < pH < 8.5$ [24]. Sample ES has a slightly acidic pH which would indicate the presence of sulfates in this stagnant water. These values are similar to those observed at the mining sites in the Pra basin in Ghana and in south-eastern of Senegal [7, 25]. The conductivity values are above the standard value of the WHO (see Table 5) for the EAM, EMS and EAV samples. This indicates that waters are mineralized [26]. The different turbidity values (see Table 1) of the samples taken at the site are high compared to the WHO standard (5 NTU). This has been also observed in the auriferous zones and the mining sites of Bonikro, Betare-Oya and Amazonia [6, 8, 27]. The waters of the site have a trouble state. At the upstream of the site, turbidity is low (38.23 NTU), because that there is no exploitation activity. Turbidity increases at the points in the site (EMS and ES) and downstream (EAV) due to excavation, river diversion and ore washing. Suspended solids are related to turbidity. It is noted that upstream EAM, the value of the SS is relatively low and below the standard WHO ($50 mg.L^{-1}$) as shown in Table 1. For the samples in the mining activity area, the SS values are well above the WHO standard (see Table 1). This is due to deforestation, the

diversion of rivers and the washing of gold [28]. Thus, this reinforces the influence of mining activity.

3.1. Heavy metals in waters

The order of the average content of heavy metals concentrations in the waters of the exploitation site is: Cu < Zn < Ni < Cd < Pb < Fe. Figure 3 presents the spatial variation of copper concentration in the water samples.

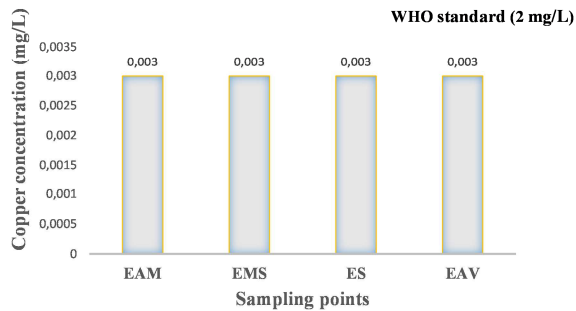


Figure 3: Spatial variation of copper concentration.

In figure 3, the concentrations do not vary according to the sampling points. They are below the WHO standard (2 mg.L^{-1}). The presence of copper in the water samples may also be due to soil leaching. The analysis of the sediment samples near the sampling points shows fairly high concentrations of copper (see Tables 2 and 3). The spatial variation of zinc concentration is presented in

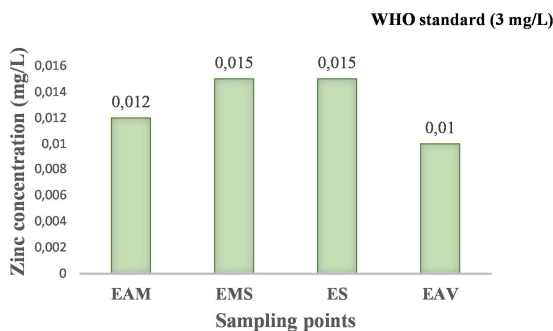


figure 4.

Figure 4: Spatial variation of zinc concentration.

Figure 4 shows that zinc concentrations are below the WHO standard which indicated a slight zinc contamination of the site water. Zinc concentrations are not elevated in gold zones [6, 8]. Natural sources of zinc in the environment are rock alteration, volcanism and vegetation [29]. In gold mining areas, zinc from sphalerite is most often associated with PbS galena [5]. The spatial variation of the concentration of nickel concentration is presented in figure 5.

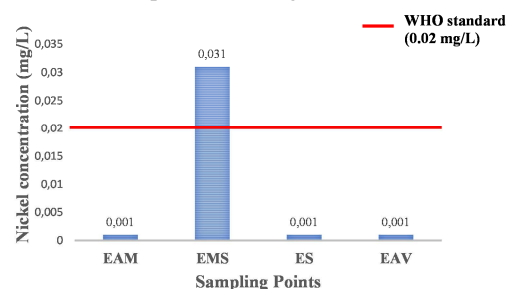
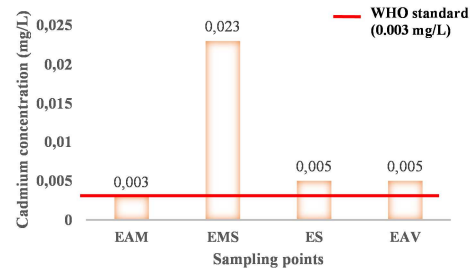


Figure 5: Spatial variation of nickel concentration.

Figure 5 shows a peak in nickel concentration at the EMS sample. This concentration is higher than the standard of WHO (0.02 mg.L^{-1}) [24]. The presence of nickel in water samples may be due to soil leaching. The analysis of the sediment sample near the sampling points shows fairly



high concentrations of nickel (see Tables 2 and 3). Figure 6 presents the spatial variation of cadmium concentration in the sampling waters.

Figure 6: Spatial variation of cadmium concentration.

In figure 6, the cadmium concentrations are above the WHO standard (0.003 mg.L^{-1}) except for the upstream sampling point. Studies at Betare-Oya on the Lomriver (East, Cameroon) has also shown that concentrations are above the standard [8]. Sphalerite is the main source of zinc and cadmium. Zinc is a trace element in sphalerite [30]. Cadmium is the most feared element because it is toxic, mobile, easily bioavailable and could either pass through the food chain through plants. Cd could come from Ni-Cd batteries used in gold mine sites [31, 32]. The spatial variation of lead concentration is presented in figure 7

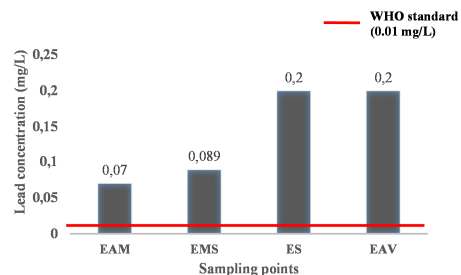


Figure 7: Spatial variation of lead concentration.

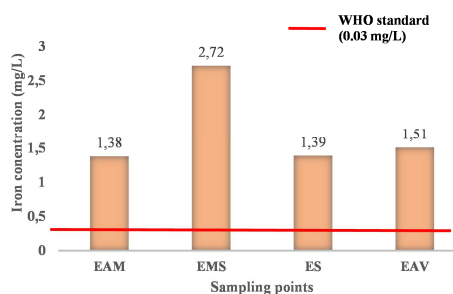
In figure 7, the lead concentrations are above the WHO standard at all the sampling points. This indicates the pollution. They are also higher than lead concentrations in the two gold zones of Bonikro and Betare-Oya [6, 8]. The main source of lead in nature is galena (PbS). The lead could be from the dissolution of this ore. In the presence of oxygen, galena can produce Pb^{2+} ions according to the equation: $\text{PbS} + 2\text{O}_2 \rightarrow \text{Pb}^{2+} + \text{SO}_4^{2-}$

.Another source of lead is the leaching of lead in equipment used for ore washing and gold mining. Lead could come from the exhaust pipes of excavators or vehicles used at mining sites [32].

In figure 8, iron concentrations are above the WHO standard (0.3 mg.L^{-1}) at all sampling points indicating the pollution. They are higher than those obtained in the Bonikro gold mine in Ivory Coast [6] and in the Betare-Oya gold zone [8]. Two origins may justify the presence of iron in the waters of the mining site. The first could be the dissolution of pyrite (FeS). In this type of mining, placer deposits come from a gold-bearing bedrock and most often this rock is accompanied by satellite minerals such as pyrite. Thus, during the deposit of gold quantities of pyrite (FeS) are present and the dissolution of this rock can contaminate the waters of the exploitation. During panning, the presence of iron oxide in the bottom of the panes is also observed. The second is the leaching of lateritic ferralitic soils with strong complexation with humic acids. Two types of soils are present in the study

area, particularly ferruginous tropical soils (the majority) and ferralitic soils [33]. Studies have also shown that iron content is generally high in West African waters due to leaching of lateritic soils [34]. The spatial variation of iron concentration is presented in figure 8.

Figure 8: Spatial variation of iron concentration.



3.2. Heavy metals in the sediments

The concentrations of heavy metals determined in sediments are compared to the WHO standard (Tables 2 and 3).

Table 2: Heavy metals content (ppm) in the sediment collected in the middle of the site (Lom).

Elements	Cu	Zn	Ni	Cd	Pb	Fe
Concentration (ppm)	8	27	39	0	17	31384
WHO standard (ppm)	30	90	50	0.35	35	/

Table 3: Heavy metal content (ppm) in the sediment taken from the stagnant water pit.

Elements	Cu	Zn	Ni	Cd	Pb	Fe
Concentration (ppm)	13	38	38	0	0	37357
WHO standard (ppm)	30	90	50	0.35	35	/

In Table 2, the concentrations are below the limits set by WHO. The concentrations of Cu, Zn and Ni are below the WHO standard are shown in Table 3. There is no lead and cadmium in the sample taken from the stagnant water pit.

3.3. Correlation matrix

The correlation between elements determined in water are presented in Table 4.

Table 4: Correlation matrix between parameters determined in water.

Variables	pH	EC	Tu	SS	Zn	Ni	Cd	Pb	Fe
pH	1	0.788	0.151	0.153	0.473	0.055	0.049	0.047	0.090
EC	0.788	1	0.282	0.283	0.188	0.196	0.153	0.410	0.224
Tu	0.151	0.282	1	1.000	0.388	0.152	0.222	0.291	0.160
SS	0.153	0.283	1.000	1	0.389	0.151	0.221	0.291	0.159
Zn	0.473	0.188	0.388	0.389	1	0.296	0.304	0.032	0.234
Ni	0.055	0.196	0.152	0.151	0.296	1	0.990	0.234	0.992
Cd	0.049	0.153	0.222	0.221	0.304	0.990	1	0.154	0.992
Pb	0.047	0.410	0.291	0.291	0.032	0.234	0.154	1	0.191
Fe	0.090	0.224	0.160	0.159	0.234	0.992	0.992	0.191	1

The relationship between heavy metals in water and sediment samples are illustrated in Table 5.

Variables	Cu	Zn	Ni	Cd	Pb	Fe
Cu	1	0.994	0.913	0.268	0.147	0.969
Zn	0.994	1	0.952	0.278	0.207	0.990
Ni	0.913	0.952	1	0.290	0.408	0.985
Cd	0.268	0.278	0.290	1	0.116	0.287
Pb	0.147	0.207	0.408	0.116	1	0.291
Fe	0.969	0.990	0.985	0.287	0.291	1

Table 5: Correlation matrix between heavy metals present in water and sediment.

In Table 4, high correlations are obtained between physical parameter, such as pH and EC (0.788). Turbidity also has a high correlation with SS (1). This correlation is attributed to physical pollution, which probably originated from the erosion of mining wastes (tailings) and discharge of wastewater from gold washing into the rivers. The values of Tu and SS are above the WHO Standard (see Table 1). For the heavy metals, the matrix correlation obtain indicated high correlations between Ni/Cd (0.990). Ni and Cd come from Ni-Cd batteries used in this gold mine site. They are also high correlations between Ni/Fe (0.992) and Cd/Fe (0.992). These elements can sometimes be derived from geological units, for example Cd exists in sphalerite (ZnS) accompanied by Zn; iron is linked to the oxidation of sulfide minerals (FeS₂ pyrite, arsenopyriteFeAsS) or rocks containing gold bearing ores [35].

The correlations are also observed between heavy metals in water and sediment samples (see Table 5). High correlations is obtained between Cu/Zn (0.994), Cu/Ni (0.913) and Cu/Fe (0.965). The origin can be from mining tailings deposited along the Lom river. Table 5 shows high correlations between Zn/Ni (0.952), Zn/Fe (0.990) and Ni/Fe (0.985). These elements are present in the composition of the sediments (see Tables 2 and 3). In the water samples, the presence of these elements can be due to soil leaching. The significant correlation observe between these parameters suggest that the origin can be the mining activity or from geological formations.

3.4. Lead adsorption by smectite

The physical characteristics of Bana smectic clay are presented in Table 6. The choice is focused on the adsorption of lead because it is the most dangerous between all heavy metals analyzed. In addition, it is a not essential metal for living organisms and its concentrations are above the WHO standard at all the sampling points compared to cadmium.

Table 6: Physical characteristics of Bana clay (smectite).

Parameters	Surface of micropores (m ² .g ⁻¹)	Surface off micropores (m ² .g ⁻¹)	Total specific surface area (m ² .g ⁻¹)	Cation Exchange Capacity(CEC)	pH
------------	--	---	--	-------------------------------	----

In Table 6, the clay sample has acid equilibrium pH. The cobalt hexane method revealed CEC value 49 meq/100 g. The specific surface area of the sample is $49.4 \text{ m}^2 \cdot \text{g}^{-1}$. This high specific surface area confirms the high smectite mineral content in this clay [17]. It is an important factor when the clay is used as adsorbent. Tables 5 and 6 show the amounts of lead adsorbed by Bana Smectite.

Electrokinetic

The electrophoretic mobility is measured by micro-electrophoresis using a Zetaphometer V by CAD Instrumentation (LEM), in 10^{-3} M NaNO_3 background in a pH range from 2 to 10. The result of the electrophoretic mobility of the smectite clay is presented in figure 9.

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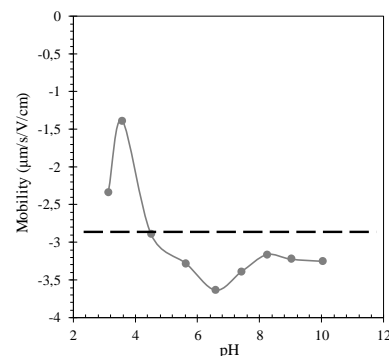


Figure X-ray

patterns of BN 61 sample.

The mineralogical composition of fraction sample less than $63 \mu\text{m}$, which is found from X-ray diffractograms, is determined by using its characteristic reflections (001). It consists of predominantly montmorillonite (70-77%) and substantial amounts of mica (5-9%), kaolinite (2-9%), quartz (1-4%), anatase (2%) and K-feldspars (1-6%) (Table 8).

Chemical analyzes

The FTIR spectra of the $< 63 \mu\text{m}$ fraction sample is shown in figure 11. The spectra for BN 61 sample shows a band at 3621 cm^{-1} , corresponding to stretching vibrations of $\text{Al}(\text{OH})$ or AlMgOH in smectite.

Figure 9: Electrophoretic mobility of the clay fraction in 10^{-3} M NaNO_3 as function of pH.

In the range of investigated pH (2 to 10), the mobility is negative as observed by the electrophoretic mobilities (EM) (See figure 9). This negative mobility indicates that the clay surface is negatively charged. A relative constant mobility between $\text{pH} = 4.5$ and $\text{pH} = 10$ is observed for an EM average value of $3.3 \mu\text{m} \cdot \text{s}^{-1} \cdot \text{V}^{-1} \cdot \text{cm}^{-1}$. No isoelectric point is observed within 4.3-10 pH range. The constant behavior between pH 4.5 and 10 is concordant with the montmorillonite behavior due to the permanent charge of this clay. The absence of an isoelectric point is usual in clays with permanent crystalline charge like smectites.

Mineralogical composition

The XRD patterns are recorded in the range $3-63^\circ 2\theta$. The result of XRD pattern of BN 61 sample clay is presented in figure 10.

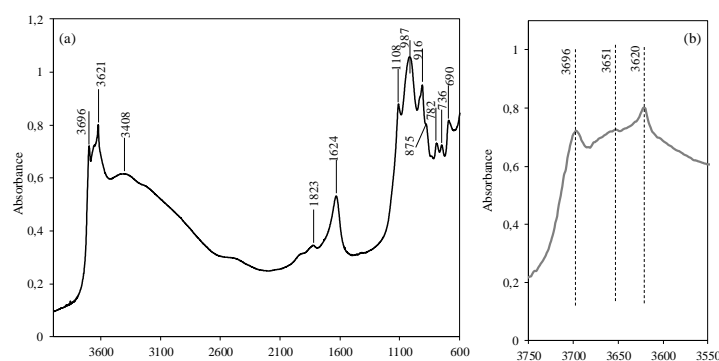
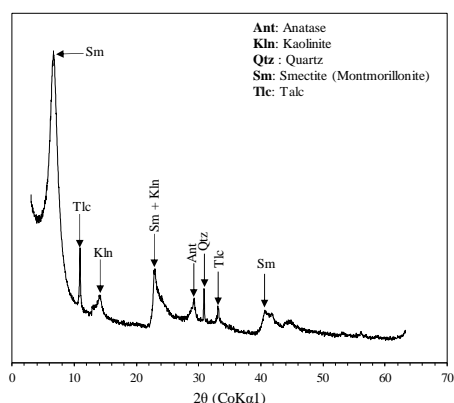


Figure 11: Infrared spectra of the clay sample (a) complete spectrum (b) OH stretching domain.

The absorbency at 3621 and 916 cm^{-1} confirms the dominant presence of dioctahedral smectite with (Al-Al-OH and Mg-OH-Al) stretching and bending bands. The weak band at 3696 cm^{-1} corresponds to Al-OH stretching vibration of kaolinite [36]. The chemical compositions of the $< 63 \mu\text{m}$ fractions are analyzed for the major elements and the results are presented in Tables 7 and 8.

Table 7: Composition in oxides (wt. %) of Banasmectic clay.

Samples	Elements											Total
	SiO ₂	Al ₂ O ₃	Fe ₂ O ₃	MnO	MgO	CaO	Na ₂ O	K ₂ O	TiO ₂	P ₂ O ₅	L.O.I	
BN 61	50.11	23.41	7.25	0.01	2.88	0.08	0.01	0.13	2.10	0.19	13.48	99.62

L.O.I: Loss on ignition.

Table 8: Mineral composition (wt. %) of Banasmectic clay.

Samples	Elements						
	Smectite	Mica	Kaolinite	Quartz	Anatase	K-feldspars	Iron oxide
BN 61	70.5	9.0	7.2	1.5	2.1	0.8	7.3

The most abundant oxides in this sample are SiO₂, Al₂O₃ and Fe₂O₃ (See Table 7) because their mineralogy is dominated by smectite. The sample exhibited a similar TiO₂ content 2.10%. The presence of this oxide is coherent with the presence of anatase. This sample is characterized by high amounts of SiO₂ (52.11%) and relative amounts of Al₂O₃ (23.41%). The Al₂O₃ content within the clay depends on the intensity of kaolinization. A higher Al₂O₃ content indicates more kaolinization of a clay sample [37]. In addition the SiO₂/Al₂O₃ is equal to 2.14. It is indicative of smectite clay with a ratio ~2.36. Bana clay are made essentially of clays such as smectite or mica (See Table 8). The Fe₂O₃ contents for this sample vary is 7.25%, indicate that the iron cation are present in the interlayer of smectite or associated with another minerals like hematite. The loss on ignition for the sample is ranged from 13.48%.

Adsorption test

The analysis is performed on the EMS and ES samples with respective lead concentrations of 0.089 mg.L⁻¹ and 0.2 mg.L⁻¹. The mass of the adsorbent is amount 0.1 to 0.5g. Experimental conditions are: volume of water = 50 mL, speed = 800 rpm, time = 1 hour and temperature = 25°C. The results of this adsorption test are presented in figure 12.

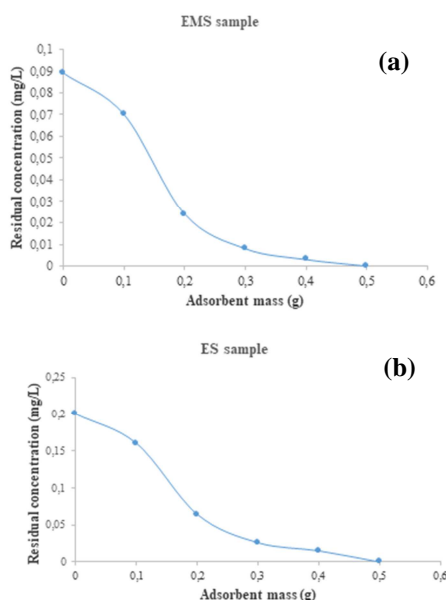


Figure 12: Residual concentrations after adsorption by Banasmectite: (a) – EMS sample; (b)- ES sample.

Samples of waters polluted with lead are subjected to adsorption in the presence of Bana clay (Ambient temperature of 26°C, agitation time 1 hour and

800turns.min⁻¹ of stirring). The quantities of lead in water samples are adsorbed. This elimination of lead by Bana clay (West, Cameroon) shows that the adsorption sites present on the surface of this material are favorable for the fixation of lead. It is also noted that 0.5g of Bana clay is sufficient for the removal of all the lead contained in 50mL of water.

4. CONCLUSION

The aim of this paper was to characterize physically and chemically the waters of the artisanal semi-mechanized gold mining site of Djouzami in order to identify probably sources of pollution and to realize a lead removal test with Bana(West, Cameroon) smectic clay. The physical parameters then the heavy metals and the test of the capacity of the clay of Bana to eliminate the lead in an aqueous medium were highlighted. The measured values of hydrogen potential, electrical conductivity, turbidity, and the suspended solids were above the World Health Organization standard for sampling points in the site and downstream of the site of operation. This pollution is mainly due to the excavation work for the extraction of gold. Zinc and copper contamination were found with very low values compared to the World Health Organization standard. At the upstream of the site, there was no pollution of heavy metals such as nickel and cadmium but higher than the standard for the sample taken in the middle of the site (Lom). For iron and lead, pollution at all sampling points was observed. The physical and chemical analyzes shown that the adsorbent material is a smectic clay. The ability of Bana clay (smectite) to remove the lead determined in the water samples was carried out. After introduction of 0.5g of Banasmectite, all lead was completely removed. Thus, the smectite of Bana which is a local material

having within its sites favourable to the fixation of the lead in aqueous medium. In the future works, it will be interesting: (i) to investigate the mechanism of adsorption that may occur in lead adsorption; (ii) using Banasmectite as adsorbent and all the correlation between the phenomenon that occurs during adsorption to determine the adsorption kinetics and isotherms.

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